

Application of Oil Palm Empty Fruit Bunch as Adsorbent: A Review

Janice May Lynn Thoe¹, Nourie Surugau¹, Harry Lye Hin Chong^{1,2,3#}

¹ Faculty of Science and Natural Resources, Universiti Malaysia Sabah, 88400 Kota Kinabalu, Sabah, MALAYSIA.

² Water Research Unit, Universiti Malaysia Sabah, 88400 Kota Kinabalu, MALAYSIA.

³ Sustainable Palm Oil Research Unit, Universiti Malaysia Sabah, 90509 Sandakan, MALAYSIA.

Corresponding author. E-Mail: hlhchong@ums.edu.my; Tel: +6088-320000; Fax: +6088-435324

ABSTRACT The abundance of oil palm wastes specifically oil palm empty fruit bunch (EFB) has possessed disposal issues that need to be tackled. Consequently, the utilisation of EFB as adsorbent for adsorbing pollutants from wastewater is a way forward. The unmodified EFB can be applied naturally but showed low adsorption capacity. The adsorption performance of EFB can be significantly improved upon modifications. This review covers the modification methods adopted to transform EFB into value-added adsorbent. Physical modifications discussed are heat pyrolysis, microwave irradiation and hydrothermal carbonisation. The output of heat pyrolysis followed by activation through oxidising gaseous and chemicals produced EFB activated carbon with high BET surface area and microporous which promotes high adsorption capacity. Besides, chemical modifications utilising acid, alkali, polymer grafting, organic and inorganic solvents provide high specificity on designing EFB adsorbent in the removal of targeted pollutants. Generally, this review serves as a guidance for researchers to move forward in searching for a simple, economic and environmental friendly technique to produce EFB based adsorbent with excellent properties and adsorption performance.

KEYWORDS: Empty fruit bunch fibre; Adsorbent; Adsorption; Modification

I Received 15 January 2019 II Accepted 21 January 2019 II In press 30 January 2019 II Online 9 February 2019 II © Transactions on Science and Technology I

INTRODUCTION

The Food and Agricultural Organisation of United Nations (FAO, 2018) had compiled the statistics on the world cultivated oil palm and the production of palm oil since year 1961. In year 2017, more than 21.09 million ha of land were cultivated with mature oil palm tree in 44 countries. The highest oil palm planting countries were Indonesia, Malaysia, Nigeria and Thailand, with the area of 9.28, 5.11, 3.04 and 0.76 million ha, respectively. The production of fresh fruit bunch (FFB) exceeded 317.57 million ton, where Indonesia (49.86%), Malaysia (32.04%), Thailand (4.59%) and Nigeria (2.44%) were the top four producing countries.

The oil palm empty fruit bunch (EFB) is generated once the oil palm fruitlets are stripped off. As much as 22% of FFB comprised of EFB (Yusoff, 2006) where 69.87 million ton of EFB was generated worldwide in year 2017. The EFB was initially applied as boiler fuel but was banned later as it gives off soot that caused air pollution to the surroundings. Since then, landfilling became the main disposal method for EFB. However, dumping EFB is also means nutrient loss which could be recycled in a gradual manner into the soil. A study by Abu Bakar et al. (2011) found that throughout the ten-year application of EFB on oil palm plantation had increased the carbon and nitrogen content in soil resulting higher FFB yield. This finding encourages the application of EFB in the plantation estates although it is time consuming for the actual outcomes to be seen. However, limitations such as labour shortage, logistical issues and oil palm pests are inevitable concerns for the management.

Another challenge for the utilisation of EFB is the high processing cost to transform EFB into value added products instead of direct utilisation. The application of EFB as a raw material for biofuel, paper production, briquette and activated carbon requires a few stages of treatment where heat, chemical, water and time is consumed at a large scale. In addition, secondary wastes maybe generated along the process of modification which eventually possesses secondary pollution if released into the environment prior to any treatment.

Oil palm EFB is a carbon rich material which is gaining new research interests in adsorption studies mainly due to its properties and abundance. In the past, the utilisation of natural EFB adsorbents has been focused on dye and heavy metal removal from wastewater. The increasing number of studies in the modifications of EFB also increased the adsorption of various pollutants from waterways. The objectives of this paper are to review (i) the methods conducted in the preparation of EFB adsorbents and (ii) the efficacy of EFB adsorbent in the removal of various pollutants from wastewater, in natural and modified forms. The published articles in various journals and proceedings from year 1997 to November 2018 are compiled and categorised based on the methods used to prepare EFB adsorbent. The literatures that utilised EFB adsorbents for the removal of various pollutants and their adsorption capacities are presented along. The efficiency of each modified adsorbent is discussed with respect to their preparation methods and the changes in physicochemical properties.

PROPERTIES OF EMPTY FRUIT BUNCH

The basic information on the characteristics of EFB is shown in Table 1. The elemental content shows that carbon is the main composition of EFB followed by oxygen, hydrogen, nitrogen and sulphur. Just like any other agricultural products, EFB is made up of lignin, hemicellulose and cellulose components. The wide range of composition for each element may be affected by the sample origin including land fertility, local weather and the soil nutrients, different analytical instruments and sample preparation prior to analysis. The high carbon and cellulose content in EFB makes it a suitable candidate to be applied as adsorbent.

Table 1. Characteristics of natural EFB

Parameters	Range	References
C (%)	40.93-68.3	Idris <i>et al.</i> (2010), Yang <i>et al.</i> (2004), Nasir <i>et al.</i> (2015)
H (%)	2.88-7.33	Wahi <i>et al.</i> (2009), Mohammed <i>et al.</i> (2012), Yang <i>et al.</i> (2004)
N (%)	< 0.1-2.18	Shariff <i>et al.</i> (2014), Abdullah & Sulaiman (2013), Lee <i>et al.</i> (2014)
S (%)	0.04-0.92	Uemura <i>et al.</i> (2011), Parshetti <i>et al.</i> (2013), Abdullah & Sulaiman (2013)
O (%)	26.4-51.78	Nasir <i>et al.</i> (2015), Idris <i>et al.</i> (2010)
Lignin (%)	10-34.37	Umikalsan <i>et al.</i> (1997), Ishola <i>et al.</i> (2014),
Hemicellulose (%)	19.5-38.8	Coral Medina <i>et al.</i> (2015); Saritpongteeraka <i>et al.</i> (2015)
Cellulose (%)	22.2-65	Mohammed <i>et al.</i> (2012), Mahjoub <i>et al.</i> (2012)
pH	7.20 – 7.80	Kavitha <i>et al.</i> (2013), Nasir <i>et al.</i> (2015)
Surface area (m ² /g)	1.48 – 28.4	Parshetti <i>et al.</i> (2012), Hidayu <i>et al.</i> (2013), Joseph <i>et al.</i> (2016); Khosravifahthany <i>et al.</i> (2013); Wirasnita <i>et al.</i> (2014), Nasir <i>et al.</i> (2015)

Besides, EFB exhibited neutral to slightly alkali character (pH 7.20 – 7.80) which is suitable for neutralising acidic soil (Kavitha *et al.*, 2013) and favouring the adsorption of cationic pollutants due to electrostatic interaction (Arshadi *et al.*, 2014). On the other hand, the Brunauer-Emmett-Teller (BET) surface area is one of the most measured characters of EFB in adsorption studies. The BET surface area of natural EFB measured ranged from 1.48 – 28.4 m²/g. The modifications of EFB are usually expected to give rise to the BET surface area.

Fourier-transform infrared (FTIR) technique played an important role in determining the surface functional groups present and adsorption mechanism involved (Setiabudi *et al.*, 2016). It is evident in Table 2 that the surface functional groups of EFB are multifunctional and varies upon modifications. The functional groups identified were hydroxyl, carboxylates, carbonyl, amides, phenol and alkyl groups among others at varied absorption wavenumbers that give rise to efficient

reduction of pollutants in targeted wastewaters. Moreover, the wavelength detected at 1738 – 1717, 1264 – 1035, and 897 – 858 cm^{-1} were evident for the presence of lignin, hemicellulose and cellulose in EFB, respectively (Daneshfozoun *et al.*, 2014a; Fatah *et al.*, 2014; Wirasnita *et al.*, 2015).

Table 2. The surface functional groups of natural EFB identified

Wavenumber (cm^{-1})	Inference	Compound	Reference
3600 – 3200	O–H stretching vibration	Hydroxyl group of hemicellulose, cellulose, lignin and the adsorbed water constituents	Haron <i>et al.</i> (2009) Johari <i>et al.</i> (2013)
2930 – 2850	C–H stretching	CH_3 group stretching vibration	Nasir <i>et al.</i> (2015) Wirasnita <i>et al.</i> (2015)
2340 – 2283	$\text{C}\equiv\text{C}$ $\text{C}\equiv\text{N}$	Alkynes Nitriles	Khosravihaftkhany <i>et al.</i> (2013)
1738 – 1717	C=O stretching	Carboxylic acid in ester group of hemicellulose or carbonyl ester of p-coumeric	Nasir <i>et al.</i> (2015) Daneshfozoun <i>et al.</i> (2014a) Wirasnita <i>et al.</i> (2015)
1680 – 1600	C=O stretching	Alkene group of lignocellulose	Nasir <i>et al.</i> (2015)
1648 – 1620	C=C C=O	Alkenes and carbonyls	Khosravihaftkhany <i>et al.</i> (2013)
1637 – 1635	N–H	Amide group	Wirasnita <i>et al.</i> (2015)
1507 – 1425	C=O vibration	Aromatic vibration in hemicellulose and lignin	Fatah <i>et al.</i> (2014)
1432 – 1319	C–H bending vibration	CH_2 bending vibrations in polysaccharides aromatic rings	Celino <i>et al.</i> (2014) Daneshfozoun <i>et al.</i> (2014a) Wirasnita <i>et al.</i> (2015)
1264 – 1035	C–O stretching Si–O stretching	Organic siloxane or silicone and anhydroglucose chains with a C–O stretch	Fatah <i>et al.</i> (2014) Wirasnita <i>et al.</i> (2015) Coates (2000)
897 – 858	C–H bending	β -glycosidic bond, out of plane bending	Celino <i>et al.</i> (2014) Fatah <i>et al.</i> (2014) Daneshfozoun <i>et al.</i> (2014a)

EMPTY FRUIT BUNCH AS ADSORBENT

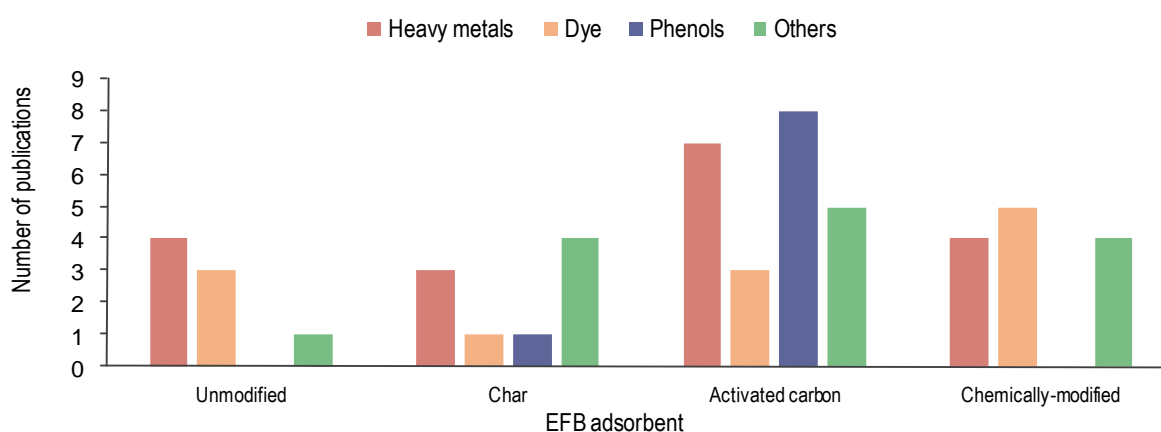


Figure 1. The types of EFB adsorbents

The utilisation of EFB as adsorbent is categorised based on modifications as shown in Figure 1. The most common adsorbates studied were heavy metals, dyes, phenols and other adsorbates include urea, oil, nutrient and pesticide and the measure of changes in colour, chemical oxygen demand (COD) and biochemical oxygen demand (BOD) in wastewater. Up to November 2018, there were 23 published articles related to EFB based activated carbon while 13 articles related to chemically modified EFB adsorbents. Publication related to the adsorption of phenolic compounds by unmodified and chemically modified EFB was not found, indicating data gap.

UNMODIFIED EMPTY FRUIT BUNCH AS ADSORBENT

Previous studies had shown natural EFB was capable of adsorbing dyes and heavy metals from aqueous solution (Table 3) (Nassar & Magdy, 1997; Rebitanim *et al.*, 2012; Joseph *et al.*, 2015; Nassar *et al.*, 2004; Salamatinia *et al.*, 2006; Khosravihaftkhany *et al.*, 2013; Daneshfozoun *et al.*, 2014a). The oxygen containing surface functional groups such as hydroxyls and carboxylic acids were claimed to be the reasons for adsorption of dyes and metal ions (Khosravihaftkhany *et al.*, 2013).

Table 3. The adsorption studies using unmodified EFB

Adsorbate	Particle size (mm)	Isotherm	Kinetic model	Q_m (mg/g)	Reference
Reactive black 5	2 – 3	Langmuir	Pseudo-second order	7.34	Joseph <i>et al.</i> (2015)
Methylene blue	0.5	Langmuir	Pseudo-second order	50.76	Rebitanim <i>et al.</i> (2012)
Basic blue	0.3	Freundlich	-	91.33	Nassar & Magdy (1997)
Basic red	0.3	Freundlich	-	180.3	Nassar & Magdy (1997)
Basic yellow	0.3	Freundlich	-	327.57	Nassar & Magdy (1997)
Cu	-	Freundlich	-	3.5945	Salamatinia <i>et al.</i> (2006)
Fe	0.215 – 0.3	Langmuir	-	1.98	Nassar <i>et al.</i> (2004)
Fe	0.2 – 1.4	Langmuir	Pseudo-second order	8.887	Khosravihaftkhany <i>et al.</i> (2013)
Mn	0.215 – 0.3	Langmuir	-	2.21	Nassar <i>et al.</i> (2004)
Pb	0.2 – 1.4	Langmuir	Pseudo-second order	0.191	Khosravihaftkhany <i>et al.</i> (2013)
Pb	0.1 – 0.2	-	-	-	Daneshfozoun <i>et al.</i> (2014a)
Zn	-	Freundlich	-	2.6021	Salamatinia <i>et al.</i> (2006)
Ammonia	-	Freundlich	Pseudo-second order	8.435	Zahrim <i>et al.</i> (2014)

The natural EFB was able to adsorb of 7.34 mg/g of reactive black 5 (Joseph *et al.*, 2015) and 50.76 mg/g methylene blue (Rebitanim *et al.*, 2012). Both studies found that the adsorption best fitted the Langmuir isotherm model and pseudo-second-order kinetic model where chemisorption was proposed as the rate determining step. Nassar and Magdy (1997) found that the adsorption of basic dyes best fitted the Freundlich isotherm where the adsorption was heterogeneous. Nassar and Magdy (1997) also suggested the utilisation of cheaper EFB as an alternative for AC and the spent EFB can be used as solid fuel.

Metal ions such iron (Fe), manganese (Mn) and lead (Pb) experienced homogeneous adsorption (Khosravihaftkhany *et al.*, 2013; Nassar *et al.*, 2004) while copper (Cu) and zinc (Zn) experienced heterogeneous adsorption (Salamatinia *et al.*, 2006) by unmodified EFB. The variation in adsorption isotherm was most probably due to the degree of leaching of water soluble organics and fatty acids from EFB which caused chemical oxygen demand (COD) introduction into wastewater during adsorption. The issue of COD can be resolved upon modifications of EFB.

PHYSICAL MODIFICATION

The physical modifications discussed in this section focused on heat pyrolysis, microwave and hydrothermal carbonisation. Sole pyrolysis step at high temperature and inert nitrogen or argon gas atmosphere produced EFB char. The continuum activation process that oxidised the EFB char is termed EFB activated carbon (AC).

Heat pyrolysis

Heat pyrolysis is the thermal conversion process under the absence of oxygen to remove volatile organic compounds yielding carbon-rich materials known as char. The varying parameters studied were temperature, heating rate, inert gas flow rate and heating time. The experimental parameters and the characteristics of EFB char are shown in Table 4.

Table 4. The conditions of pyrolysis and the characteristics of EFB char

Pyrolysis				BET	Average	Adsorbate	Q _m (mg/g)	Reference
Temp. (°C)	Heating rate (°C/min)	N ₂ flow rate (L/min)	Time (h)	surface area (m ² /g)	pore diameter (nm)			
-	-	-	-	255.77	2.232	Methylene blue	-	Foo & Hameed (2011)
-	-	-	-	290.35	-	Phenol	-	Yap <i>et al.</i> (2005)
-	-	-	-	1.89	2.4818	Cu	49.4	Samsuri <i>et al.</i> (2014)
-	-	-	-	1.89	2.4818	Pb	58.8	Samsuri <i>et al.</i> (2014)
-	-	-	-	1.89	2.4818	Zn	45.7	Samsuri <i>et al.</i> (2014)
300	30	-	0.33	4.54	-	-	-	Sukiran <i>et al.</i> (2011)
300	3	0.1	1	1.46	-	Imazapyr	-	Yavari <i>et al.</i> (2017)
300	-	-	-	44.38	2.884	Zn	-	Zamani <i>et al.</i> (2017)
300-350	-	-	-	46.32	3.85	As	0.42	Sari <i>et al.</i> (2014)
300-350	-	-	-	46.32	3.85	Cd	15.15	Sari <i>et al.</i> (2014)
400	30	-	0.33	5.76	-	-	-	Sukiran <i>et al.</i> (2011)
500	30	-	0.33	4.85	-	-	-	Sukiran <i>et al.</i> (2011)
500	-	-	1	9.09	-	-	-	Wirasnita <i>et al.</i> (2015)
550	5	-	1	11.12	-	-	-	Shariff <i>et al.</i> (2014)
600	30	-	0.33	3.95	-	-	-	Sukiran <i>et al.</i> (2011)
615	8	0.15	2.13	421.26	1.441	Zn	-	Zamani <i>et al.</i> (2017)
700	30	-	0.33	3.34	-	-	-	Sukiran <i>et al.</i> (2011)

According to Table 4, the EFB char produced at temperature 300 – 700°C resulted in low BET surface area (3.34 – 5.76 m²/g) showed not much significance when compared to natural EFB (Sukiran *et al.*, 2011). The high heating rate of 30°C/min was the reason where the decomposition was not able to occur uniformly resulted in low BET surface area measured. The BET surface area can be improved where the pyrolysis conducted at low heating rate of 3 – 8°C/min with prolonged reaction time could eventually achieved BET surface area of 421.26 m²/g (Zamani *et al.*, 2017). Despite of the low BET surface area of EFB char, the average pore diameter indicated the development of mesopores and micropores. Hence, activation of EFB char was conducted by researchers to improve the BET surface area, development of mesopores and micropores to achieve high adsorption capacity.

Char Activation

The activation process in developing EFB AC is crucial. Activation process can be conducted through gaseous activation or by chemical means. The changes in BET surface area and the adsorption performance is discussed.

Gaseous activation

The EFB char was produced when heat pyrolysis reached the targeted temperature. Activation step was initiated when the oxidising gas such as air, carbon dioxide (CO₂) or steam were allowed to flow in replacing nitrogen gas, held for a period of time at temperature 700 – 1000°C to further encouraged surface area and micropores development (Table 5).

Table 5. The EFB AC prepared through gaseous activation

Pyrolysis		Activation		BET surface area (m ² /g)	Average pore diameter (nm)	Adsorbate	Q _m (mg/g)	Reference
Temp. (°C)	Heating rate (°C/min)	Agent	Temp (°C)					
-	-	Air	800	-	-	2,4-C ₆ H ₄ Cl ₂ O	27.25	Alam <i>et al.</i> (2007)
-	-	Air	1000	-	-	Zn	-	Alam <i>et al.</i> (2008)
700	8	CO ₂	900	-	-	Phenol	66.67	Arshad <i>et al.</i> (2012)
900	23	CO ₂	900	345.10	-	Phenol	-	Alam <i>et al.</i> (2009)
500	-	Steam	765	720	1.889	-	-	Hidayu <i>et al.</i> (2013)
900	20	Steam	900	635.16	4	Cd	-	Ma'an <i>et al.</i> (2011)
900	20	Steam	900	886.2	3.54	Mn	0.0315	Amosa (2015)
900	20	Steam	900	886.2	3.54	H ₂ S	0.0083	Amosa (2015)
950	20	Steam	900	886.2	3.54	COD	15.87	Amosa <i>et al.</i> (2016)
900	23	Steam	900	-	-	Hg	0.00088	Kabbashi <i>et al.</i> (2011)

Steam activation is a better activating agent in preparing EFB AC. The high BET surface area (635.16 – 886.2 m²/g), mesopore and micropore development after steam activation is due to the smaller particles that are able to diffuse and react within the carbon matrix at higher reaction rate compared to air and CO₂ (Hashemipour *et al.*, 2009). In addition, the surface functional groups were found to disappear after steam activation (Hidayu *et al.*, 2013). The disappearance of surface functional groups during steam activation resulted in low adsorption capacity of metal ions due to the absence of electrostatic attraction. However, steam activation under high activating temperature at 765 – 900°C required high energy consumption and less cost effective for large scale production.

On the other hand, the activation process can be optimised using two-level full factorial design. Alam *et al.* (2009) and Ma'an *et al.* (2011) adopted full factorial design in the optimisation of the activation temperature, time and gaseous flow rate in producing EFB AC. The steam activated EFB AC after the optimised activation at 900C, at 2.0 mL/min steam flow rate for 15 min were able to develop EFB AC with BET surface area of 635.16 m²/g (Alam *et al.*, 2009). Moreover, Kadir *et al.* (2014) conducted steam activation on EFB AC using response surface methodology after the study by Hidayu *et al.* (2013) discovered that experimental value of BET surface area, 720 m²/g was agreed to the predicted value 717.6 m²/g through the optimised conditions of the statistical analysis. The optimisation of EFB AC using full factorial experimental design is not a favourable approach due to increasing number of experimental runs especially when the number of parameters studied exceeded four (Rashidi & Yusup, 2017).

Chemical activation

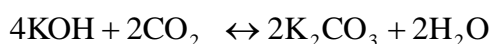
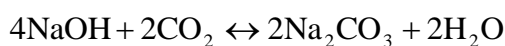
The chemical activation can occur before and after the pyrolysis of EFB. Table 6 summarises the chemical activation of EFB after pyrolysis. Chemical activation is capable of increasing the surface area and pore volume by removing tar and volatile matter that clogged the internal pores (Wahi *et al.*, 2009; Wirasnita *et al.*, 2015). The potassium hydroxide (KOH) and sodium hydroxide (NaOH) are strong oxidising agents that increased the BET surface area and promoted mesopores and micropores development (Wahi *et al.*, 2009; Yap *et al.*, 2005).

Chemical activation can be conducted through dry and wet impregnation. Dry impregnation refers to the mixing of dry KOH pellets with EFB char at fixed impregnation ratio followed by pyrolysis or carbonisation. Strong oxidising KOH coupled with CO₂ gas activation produced high BET surface area of 1141 m²/g EFB AC (Hameed *et al.*, 2009; Tan & Hameed, 2010; Tan *et al.*, 2009). The role of KOH as catalyst accelerated the rate of decomposition through dehydration and the formation of potassium carbonate (Guo & Lua, 2002).

Table 6. The chemical activated EFB AC

Pyrolysis		Activation			BET surface area (m ² /g)	Adsorbate	Q _m (mg/g)	Reference
Temp (°C)	Heating rate (°C/min)	Agent	Temp (°C)	Time (h)				
700	10	CO ₂ 74% w/w KOH	814	1.9	-	2,4,6-C ₆ H ₂ Cl ₃ OH	500	Tan <i>et al.</i> (2009)
700	10	CO ₂ 74% w/w KOH	700	1.9	1141	2,4,6-C ₆ H ₂ Cl ₃ OH	-	Hameed <i>et al.</i> (2009)
700	10	CO ₂ 74% w/w KOH	844	1.8	-	Methylene blue	416.07	Tan & Hameed (2010)
800	5	71% w/w KOH	-	2.33	820	Acid Red 1	-	Auta <i>et al.</i> (2012)
400	-	20% w/v NaOH	60	2	379.37	Hg	52.67	Wahi <i>et al.</i> (2009)
400	-	N ₂	700	1	-	-	-	-
400	-	20% w/v NaOH	60	2	379.37	Pb	48.96	Wahi <i>et al.</i> (2009)
400	-	N ₂	700	1	-	-	-	-
400	-	20% w/v NaOH	60	2	379.37	Cu	0.84	Wahi <i>et al.</i> (2009)
400	-	N ₂	700	1	-	-	-	-
-	-	1000 mg/L FeCl ₃	-	24	-	As(III)	31.4	Samsuri <i>et al.</i> (2013)
-	-	1000 mg/L FeCl ₃	-	24	-	As(V)	15.2	Samsuri <i>et al.</i> (2013)

Wet impregnation is the immersion of EFB char in dissolved oxidising chemical prepared at desired concentration. Wahi *et al.* (2009) soaked the EFB char in 20% (w/v) of NaOH solution (equivalent to 5 M NaOH) for 2 h as the activation step followed by second carbonisation at higher temperature at 700°C under inert atmosphere. The experimental results showed that dry mixing produced EFB AC with BET surface area measuring 378.55 – 1141 m²/g while wet mixing produced EFB AC with 379.37 m²/g. The advantage of dry impregnation is cross contamination can be prevented while wet impregnation utilised water as medium is easily exposed to the diffusion of CO₂ from the atmosphere, causing interruptions through the chemical reactions shown (Lillo-Rodenas *et al.*, 2003):



Besides, the impregnation of transition metal compounds, iron chloride (FeCl₃) onto EFB char increased the adsorption capacity of As(III) and As(V) by coordination and surface complexation of the ferric compounds coated into the EFB char (Samsuri *et al.*, 2013). The grafting of metal compounds onto EFB AC is more economically feasible as second carbonisation step can be skipped. However, the residual chemicals in the EFB AC caused secondary pollution and possessed toxicity towards the living organisms as it is more difficult to remove the residual chemicals within micropores. This problem can be overcome by rinsing with diluted acid which aimed to neutralise and remove the residual chemicals. For example, Wahi *et al.* (2009) utilised 5 M HCl to neutralise the excess NaOH at the end of modification to lower down the pH value in order to achieve more accurate adsorption results, especially the adsorption of metal ions is highly pH dependent.

On the other hand, the EFB treated with chemicals followed by heat pyrolysis is summarised in Table 7. Chemicals such as sulphuric acid (H₂SO₄), phosphoric acid (H₃PO₄), KOH and zinc chloride (ZnCl₂) were used to treat natural EFB prior heat pyrolysis and activation. Lee *et al.* (2014) and Ooi *et al.* (2017) treated EFB with dehydrating H₂SO₄ followed by pyrolysis and activation produced high BET surface area and micropores EFB. The removal of acid-soluble compounds within EFB allowed to the pyrolysis and activation to take place at higher degradation rate during pyrolysis where the volatilisation of organic compounds was facilitated thus forming more pores within the EFB AC.

Moreover, the application of concentrated phosphoric acid (H_3PO_4) followed by pyrolysis produced $1031.5 \text{ m}^2/\text{g}$ BET surface area EFB AC (Shaarani *et al.*, 2010). This indicated pyrolysis after chemical treatment was sufficient to develop EFB AC and further activation became unnecessary. Similar study was further enhanced with 10% (w/w) ammonia which then increased the maximum adsorption capacity of 2,4- $C_6H_3Cl_2OH$ from 232.56 mg/g to 285.71 mg/g due to the increase in basic functional groups of ammonia (Shaarani *et al.*, 2011). The surface functional groups analysis indicated the presence cyclic amides and amine group where the positively charged new nitrogen surface complexes improved the 2,4- $C_6H_3Cl_2OH$ uptake.

Additionally, the EFB mixed with KOH pellets followed by pyrolysis had BET surface area of $663 \text{ m}^2/\text{g}$ where the external surface appeared to be porous and irregular (Abdul Khalil *et al.*, 2013). The adsorption of Rhodamine B dye was studied through fixed bed column conducted by Auta (2012) utilising KOH treated EFB AC. The maximum adsorption capacity 69.86 mg/g of Rhodamine B dye was obtained at 200 mg/L of influent dye concentration, 10 cm bed depth and 15 mL/min of solution flow rate. The results obtained were found to fit the Thomas model adequately in describing the breakthrough behaviour of the adsorption process.

Table 7. Chemical activation followed by heat pyrolysis

Activation			Pyrolysis + Activation			BET	Adsorbate	Q_m (mg/g)	Reference
Agent	Tem p (°C)	Time (h)	Agent	Tem p (°C)	Flow rate (L/min)	surface area (m^2/g)			
60% w/w H_2SO_4	-	-	N_2	400	0.1	-	Urea	-	Ooi <i>et al.</i> (2017)
43% w/w H_2SO_4	-	-	N_2	400	0.1	990	-	-	Lee <i>et al.</i> (2014)
conc. H_3PO_4	120	24	-	800	-	850.11	Cu	333.33	Nwabanne & Igbokwe (2012)
20% w/w H_3PO_4	RT	Overnigh t	N_2	450	0.15	1031.5	2,4- $C_6H_3Cl_2OH$	232.56	Shaarani & Hameed (2010)
20% w/w H_3PO_4	RT	48	N_2	450	0.15	-	2,4- $C_6H_3Cl_2OH$	285.71	Shaarani & Hameed (2011)
10% w/w NH_3	-	-	N_2	-	0.15	-	Rhodamine B	69.86	Auta (2012)
67% w/w KOH	-	-	N_2	375	1	290.35	Phenols	-	Yap <i>et al.</i> (2005)
0% w/w KOH	-	-	N_2	700	1	378.55	Phenols	-	Yap <i>et al.</i> (2005)
10% w/w KOH	RT	24	N_2	375	1	417.70	Phenols	90.09	Yap <i>et al.</i> (2005)
30% w/w KOH	RT	24	N_2	700	1	415.84	Phenols	91.74	Yap <i>et al.</i> (2005)
50% w/w KOH	RT	24	N_2	375	1	431.52	Phenols	89.29	Yap <i>et al.</i> (2005)
70% w/w KOH	RT	24	N_2	700	1	663	-	-	Abdul Khalil <i>et al.</i> (2013)
75% w/w KOH	85– 90	24	-	800	-	663	-	-	Abdul Khalil <i>et al.</i> (2013)
10% w/v $ZnCl_2$	-	24	N_2	500	2.5	86.62	Bisphenol A	41.98	Wirasnita <i>et al.</i> (2014)

Strong dehydrating agent $ZnCl_2$ is known to promote charring and eliminate organic compounds prior to pyrolysis was able to increase the BET surface area from $9.09 \text{ m}^2/\text{g}$ to $86.62 \text{ m}^2/\text{g}$ (Wirasnita *et al.*, 2015). The maximum monolayer adsorption capacity of bisphenol A was 41.98 mg/g which represented by Langmuir isotherm. Despite the lower BET surface area was accounted, the proposed adsorption process of bisphenol A was chemisorption where the carbon-carbon double bonds, hydrogen bonding between carboxyl and/or hydroxyl groups and the electron donor-acceptor complexes were involved (Wirasnita *et al.*, 2014).

Microwave irradiation

Microwave provides energy to entire samples via a dielectric heating, whereby heat is transferred from the central part towards the outer surface producing more uniformly modified EFB

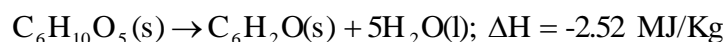
AC (Ahmad *et al.*, 2018). Microwave irradiation also serves as an alternative for pyrolysis which saves time and energy meanwhile providing instant control and accelerate the rate of reaction (Omar *et al.*, 2011; Puligundla *et al.*, 2016).

Mubarak *et al.* (2014) utilised microwave in heating EFB in ferric chloride hexahydrate instead of pyrolysis. The modifications were optimised by using response surface methodology. The optimised conditions identified were irradiation of EFB in FeCl₃ with impregnation ratio 0.5 under 900 W microwave power for 20 min. The FeCl₃ impregnated EFB have high BET surface area of 890 m²/g where the highest adsorption capacity of 265 mg/g methylene blue was adsorbed. The advantages of microwave technique is simple, safe to operate and time saving in the modifications step while yielding quality EFB adsorbent.

Foo and Hameed (2011) studied EFB AC impregnated with KOH followed by microwave irradiation at 2.45 GHz. Microwave irradiation improved the BET surface area from 255.77 m²/g to 807.54 m²/g. The adsorption of methylene blue had achieved 344.83 mg/g of maximum monolayer adsorption capacity. However, the improvement in terms of maximum monolayer adsorption capacity was not significant as the KOH treated EFB AC without microwave irradiation was able to remove 357.14 mg/g of methylene blue (Tan & Hameed, 2010).

Hydrothermal carbonisation

Hydrothermal carbonisation is a process that converts lignocellulosic materials into solid carbon with low oxygen to carbon ratio (Jamari & Howse, 2012). This process is gaining attention in the recent years due to its low operating temperature, where low energy also implies reduction in operating costs. Moreover, hydrothermal carbonisation is an exothermic process that is more environmental friendly as no gases is evolved throughout the process. The output of hydrothermal carbonisation is carbon-rich materials and wastewater which could be described as follow (Jamari & Howse, 2012).



The EFB was suspended in water under pressure controlled environment and temperature lower than pyrolysis (150 – 350°C) for a period of time (Parshetti *et al.*, 2013). The mechanisms involved including dehydration, decarboxylation, hydrolysis, aromatisation and condensation polymerisation that converted EFB into high carbon content adsorbent while maintaining the carbohydrate structure of EFB (Funke & Ziegler, 2010; Jamari & Howse, 2012). The end product was often termed EFB hydro-char. However, the production of EFB hydrochar was modified to harness energy. Hence, data gap is identified where the hydrothermally treated EFB is yet to explore its adsorption performance.

CHEMICAL MODIFICATIONS

Chemical modifications on EFB adsorbents increase the adsorption efficiency of various pollutants. The chemicals used in modifying EFB including acids, alkalis, polymers, organic and inorganic solvents. The studies EFB are discussed in this section.

Acid and alkali

The acid and alkali modified EFB adsorbents and the adsorption studies are summarised in Table 8. Acids are widely utilised as the modifying agent in developing EFB adsorbents. The acidic reactions involved hydrolysis which partially cleaves off the lignin and hemicellulose content in EFB and to remove the acid soluble impurities on the surface of EFB (Lenihan *et al.*, 2010).

Nasir *et al.* (2015) treated EFB with 0.1 M hydrochloric acid (HCl) discovered a decreased in BET surface area from 28.4 m²/g to 8 m²/g which can be explained by the collapse of the EFB pore channels. However, the reduction in BET surface area did not affect the removal performance of methylene blue and Cu ions. The adsorption of methylene blue by HCl modified EFB was homogeneous while Cu ions was heterogeneous, each represented by Langmuir and Freundlich isotherm model, respectively. The removal was attributed to the presence of carbonyl, hydroxyl, carboxylic, alkanes, alkyls, and esters in the HCl modified EFB instead of the surface area (Nasir *et al.*, 2015).

Table 8. The acid and alkali modified EFB adsorbent

Modification	Adsorption studies						Reference
	Chemical	Tem (°C)	Time (h)	Adsorbate	Isotherm	Kinetic	
conc H ₂ SO ₄	150	24	Methylene blue	-	-	-	Saad <i>et al.</i> (2007)
64% (w/w) H ₂ SO ₄	45	45	Methylene blue	Langmuir	P2 IPD	144.93	Shanmugarajah <i>et al.</i> (2019)
0.1 M HCl	-	Overnight	Methylene blue	Langmuir	-	32.944	Nasir <i>et al.</i> (2015)
0.1 M HCl	-	Overnight	Cu	Freundlich	-	4.766	Nasir <i>et al.</i> (2015)
0.6 M Citric acid	RT	0.5	Methylene blue	Langmuir	P2 IPD	103.1	Sajab <i>et al.</i> (2013)
0.6 M Citric acid	RT	0.5	Cu	Langmuir	-	126.85	Sajab <i>et al.</i> (2017)
			Ni			7.864	
10% (w/w) Acetic acid	100	2	Mn	Dubinini- Radushkevich	P2	0.003	Daneshfozoun <i>et al.</i> (2014b)
			Ni			0.003	
			Cu			0.005	
1.25 M NaOH	RT	1 12 72	Ammonia	-	-	-	Zahrim <i>et al.</i> , 2014

RT: Room temperature

P2: Pseudo-second-order

IPD: Intraparticle diffusion

Shanmugarajah *et al.* (2019) isolated nanocrystalline cellulose from EFB through 4% (w/w) NaOH treatment, bleaching with acetate buffer and sodium chlorite, followed by hydrolysis using 64% (w/w) sulphuric acid to study its methylene blue adsorption behaviour. The modification was successful where 144.93 mg/g of maximum monolayer adsorption capacity was determined through the fitting into Langmuir isotherm. The adsorption was described to be homogenous with chemisorption as the rate limiting step through the fitting of pseudo-second-order kinetics.

Alkalis NaOH and KOH are the common chemical used in the industries in treating lignocellulosic materials. Alkali is capable of breaking the intramolecular and intermolecular hydrogen bonds between hydroxyl groups (-OH) of cellulose, hemicellulose and lignin moieties, leading to the defibrillation of fibres (Chowdhury *et al.*, 2013). This process also known as mercerisation in the industrial process. Ibrahim *et al.* (2009) and Ibrahim *et al.* (2010) precipitated lignin from EFB using 20% (w/v) NaOH at 170°C which were able to remove 7.94 mg/g and 46.72 mg/g of Cu(II) and Pb(II) ions, respectively.

On the other hand, 0.10 – 5.0 M NaOH was first applied to initiate modifications and to enable the grafting of polyethyleneimine (Sajab *et al.*, 2013) and surfactant cetyltrimethylammonium bromide (CTAB) (Danish *et al.*, 2015) onto EFB to adsorb phenol red and methyl orange, respectively. The reason for NaOH application was to partially removed lignin and hemicellulose in EFB, altered the fibre structure enabling the penetration of chemicals and reduce the reaction time.

Polymer grafting

Polymer grafting on EFB aimed to increase the variety of pollutants to be adsorbed and to enhance adsorption capacity of pollutants (Table 9). The removal mechanisms by polymer grafted EFB adsorbents removed pollutants by chelation and coordination instead of ion exchange and surface adsorption by functional groups.

The grafting technique was complex where a few chemicals were needed and involved multiple steps. First, Haron *et al.* (2009) initiated the modifications using hydrogen peroxide followed by the addition methyl acrylate monomer for polymerisation yielding poly(methyl acrylate) on EFB. The intermediate was then treated with hydroxylamine in methanolic solution at pH 13 producing poly(hydroxamic acid) grafted EFB. The grafted EFB followed the Langmuir isotherm and pseudo-second-order model where the maximum monolayer adsorption capacity of Cu(II) ions achieved was 74.1 mg/g. The proposed adsorption mechanism was chemisorption, spontaneous and exothermic.

Table 9. Polymer grafting onto EFB adsorbent

Modification		Adsorption studies						Reference
Agent		Temp. (°C)	Time (h)	Adsorbate	Isotherm	Kinetic	Q _m (mg/g)	
5% Polyethyleneimine	(w/v)	65	6	Phenol red	Freundlich	P2 IPD	156.7	Sajab <i>et al.</i> (2013)
5% Polyethyleneimine	(w/v)	65	6	BOD TOC Color	-	P2 IPD	-	Sajab <i>et al.</i> (2014)
5% Polyethyleneimine	(w/v)	65	6	Mo(VII) As(V)	Freundlich Langmuir	IPD	187.77 73.42	Sajab <i>et al.</i> (2017)
Poly(Hydroxamic acid)		75	2	Cu	Langmuir	P2	74.1	Haron <i>et al.</i> (2009)
Poly(ethyl Hydrazine)		Reflux	4	Ni	Langmuir	P2	42.19	Johari <i>et al.</i> (2013)

P2: Pseudo-second-order

IPD: Intraparticle diffusion

Besides, similar study by Johari *et al.* (2013) which also engrafted poly(methyl acrylate) onto EFB as the intermediate followed by reflux in hydrazine hydrate solution with 15% (v/v) ethanol for 4 h yielding final product poly(ethyl hydrazine) grafted EFB. The chelation of Ni ions by poly(ethyl hydrazine) was proposed to be adsorption mechanism where stable complex metal ion was formed. The adsorption isotherm fitted well with Langmuir isotherm and followed the pseudo-second-order kinetics. The adsorption occurred was endothermic, spontaneous and random at the solid-solution interface. Despite of the promising adsorption capacity of the polymer grafter EFB adsorbent, high capital investment and energy consumption for large scale production are a step-back for the industries. Furthermore, the remaining reagents and solutions after modifications turned into wastewater that requires treatment before disposal.

Organic solvent and inorganic solvent

Organic and inorganic solvents are utilised as modifying agents (Table 10). Modifying agents such as acetic anhydride (Asadpour *et al.*, 2016) and trimethylchlorosilane (Rattanawong *et al.*, 2007) increased the hydrophobicity of EFB which increased the adsorption capacity of oil adsorption. Both studies have a common step of utilising NaOH to eliminate lignin and hemicellulose in EFB allowing acetylation and silylation to take place. The substitution of acetyl group and silyl groups increased the esters groups while eliminated the hydroxyl groups which benefited the oil adsorption. The oil adsorption followed the homogenous adsorption evident by the best fitting of Langmuir isotherm.

Table 10. Organic and inorganic solvents as modifying agent

Modification			Adsorption studies				Reference
Agent	Temp (°C)	Time (h)	Adsorbate	Isotherm	Kinetic	Q _m (mg/g)	
1% (v/v) trimethylchlorosilane	RT	3	Oil	Langmuir	-	0.7594	Rattanawong <i>et al.</i> (2007)
Acetic anhydride	120	4	Tapis crude oil Arabian crude oil	Langmuir	P1 P2	10000 10000	Asadpour <i>et al.</i> (2016)
1% (w/v) Formaldehyde	150	24	Methylene blue	-	-	-	Saad <i>et al.</i> (2007)
1% (v/v) Cetyltrimethylammonium bromide	-	Overnight	Methyl orange	Langmuir	P1 P2 IPD	18.1	Danish <i>et al.</i> (2015)
10% (w/w) EDTA	100	2	Mn Ni Cu	Dubinin-Radushkevich	P1 P2	0.004 0.003 0.007	Daneshfozoun <i>et al.</i> (2014b)

RT: Room temperature

P1: Pseudo-first order

P2: Pseudo-second order

The adsorption methylene blue by 1% (w/v) formaldehyde modified EFB was reported to achieve 96.4% of removal performance at initial concentration of 50 mg/L (Saad *et al.*, 2007). However, the adsorption isotherm and kinetic were not discussed in this article. Besides, the adsorption of methyl orange was conducted by EFB modified using 1% (v/v) cetyltrimethylammonium bromide (Danish *et al.*, 2015). The adsorption followed the Langmuir isotherm where maximum monolayer adsorption was 18.1 mg/g. The methyl orange adsorption was exothermic and spontaneous which indicated no energy barrier to initiate the adsorption process.

Daneshfozoun *et al.* (2014b) compared the adsorption performance of Mn, Ni and Cu by EFB treated with 10% (v/v) ethylenediaminetetraacetic acid (EDTA) and 10% (w/w) acetic acid. The EDTA contains four carboxylic acid groups while acetic acid contains one carboxylic acid group. The EDTA-modified EFB showed better adsorption performance due to the excess carboxylic acid groups. The adsorption systems were well fitted in Dubinin-Radushkevich isotherm where the apparent energy suggested physisorption mechanism.

The organic and inorganic solvents removed and/or added surface functional groups improving the surface tension between liquids making EFB more susceptible for adsorption of pollutants, specifically the adsorption of oil. The utilisation of organic and inorganic solvents can be applied to modified adsorbents in treating the targeted pollutants.

FUTURE PROSPECTS

Converting EFB into value-added adsorbents is a way to solve the disposal issue and substituting the conventional adsorbents. The utilisation of EFB as an adsorbent has gained recognitions owing to its abundance, relatively low cost and rich in lignin, cellulose and hemicellulose. Physical and chemical modifications on EFB are able to transformed EFB into value-added adsorbents with high adsorption capacity. The target application of adsorbents should be determined first prior to conduct the modifications. Thus, EFB should be explored further in terms of high BET surface area, porosity and adsorption rate. The production costs can be overcome if chemical such as acids and alkalis are coupled with physical modification in developing EFB

adsorbents. Nonetheless, modifications that are more environmental friendly, able to produce adsorbents with excellent physicochemical properties and adsorption performance is yet to be explored and developed, especially for pilot scale to industrial applications.

AKNOWLEDGEMENT

The authors would like to express their gratitude to the Malaysian Ministry of Higher Education under the Universiti Malaysia Sabah Research Priority Area Scheme (Project SBK0415-018), Postgraduate Research Grant Scheme (Project GUG0070-SG-2/2016) for the technical and financial support and the MyBrain15 scholarship for the first author.

REFERENCES

- [1] Abdul Khalil, H.P.S., Jawaid, M., Firoozian, P., Rashid, U., Islam, A. and Akil, H.M. (2013). Activated carbon from various agricultural wastes by chemical activation with KOH: Preparation and characterisation. *Journal of Biobased Materials and Bioenergy*, **7**, 1 – 7.
- [2] Abdullah, N. and Sulaiman, F. (2013). The thermal properties of the washed empty fruit bunches of oil palm. *Journal of Physical Science*, **24**(2), 117 – 137.
- [3] Abu Bakar, R., Darus, S.Z., Kulaseharan, S. and Jamaluddin, N. (2011). Effects of ten year application of empty fruit bunches in an oil palm plantation on soil chemical properties. *Nutrient in Agroecosystems*, **89**, 341 – 349.
- [4] Ahmad, M.I., Rizman, Z.I., Rasat, M.S.M., Alauddin, Z.A.Z., Soid, S.N.M., Aziz, M.S.A., Mohamed, M., Amini, M.H.M. and Amin, M.F.M. (2018). The effect of torrefaction on oil palm empty fruit bunch properties using microwave irradiation. *Journal of Fundamental and Applied Sciences*, **9**, 924 – 940.
- [5] Alam, M.Z., Ameen, E.S., Muyibi, S.A. and Kabbashi, N.A. (2009). The factors affecting the performance of activated carbon prepared from oil palm empty fruit bunches for adsorption of phenol. *Chemical Engineering Journal*, **155**, 191 – 198.
- [6] Alam, M.Z., Muyibi, S.A. and Kamaldin, N. (2008). Production of activated carbon form oil palm empty fruit bunches for removal of zinc. *Twelfth International Water Technology Conference (IWTC 12)*, 373 – 383.
- [7] Alam, M.Z., Muyibi, S.A., Mansor, M.F. and Wahid, R. (2007). Activated carbons derived from oil palm empty-fruit bunches: Application to environmental problems. *Journal of Environmental Sciences*, **19**, 103 – 108.
- [8] Amosa, M. K. (2015). Process optimization of Mn and H₂S removals from POME using an enhanced empty fruit bunch (EFB)-based adsorbent produced by pyrolysis. *Environmental Nanotechnology, Monitoring and Management*, **4**, 93 – 106.
- [9] Amosa, M. K., Jami, M. S., & Alkhatib, M. F. R. (2016). Electrostatic Biosorption of COD, Mn and H₂S on EFB-Based Activated Carbon Produced through Steam Pyrolysis: An Analysis Based on Surface Chemistry, Equilibria and Kinetics. *Waste and Biomass Valorisation*, **7**(1), 109 – 124.
- [10] Arshad, S.H.M., Aziz, A.A., Ngadi, N. and Amin, N.S. (2012). Phenol adsorption by activated carbon of different fibre size derived from empty fruit bunches. *Journal of Oil Palm Research*, **24**, 1524 – 1532.
- [11] Arshadi, M., Amiri, M.J., Mousavi, S. (2014). Kinetic, equilibrium and thermodynamic investigations of Ni(II), Cd(II), Cu(II) and Co(II) adsorption on barley straw ash. *Water Resources and Industry*, **6**, 1 – 17.

- [12] Asadpour, R., Sapari, N.B.M., Isa, M.H. and Kakooei, S. (2016). Acetylation of oil palm empty fruit bunch fibre as an adsorbent for removal of crude oil. *Environmental Science and Pollution Research*, **23**(12), 11740 – 11750.
- [13] Auta, M. (2012). Fixed bed adsorption studies of rhodamine B dye using oil palm empty fruits bunch activated carbon. *Journal of Engineering Research and Studies*, **3**(3), 3 – 6.
- [14] Auta, M., Jibril, M., Tamuno, P.B.L. and Audu, A.A. (2012). Preparation of activated carbon from oil palm fruit bunch for the adsorption of Acid Red 1 using optimised response surface methodology. *International Journal of Engineering Research and Application (IJERA)*, **2**(3), 1805 – 1815.
- [15] Celino, A., Goncalves, O., Jacquemin, F., Freour, S. (2014). Qualitative and quantitative assessment of water sorption in natural fibres using ATR-FTIR spectroscopy. *Carbohydrate Polymers*, **101**, 163 – 170.
- [16] Chowdhury, M.N.K., Beg, M.D.H., Khan, M.R. and Mina, M.F. (2013). Modification of oil palm empty fruit bunch fibres by nanoparticle impregnation and alkali treatment. *Cellulose*, **20**, 1477 – 1490.
- [17] Coral Medina, J.D., Woiciechowski, A., Filho, A.Z., Nosedá, M.D., Kaur, B.S. and Soccol, C.R. (2015). Lignin preparation from oil palm empty fruit bunches by sequential acid/alkaline treatment – A Biorefinery approach. *Bioresource Technology*, **194**, 172 – 178.
- [18] Daneshfozoun, S., Abdullah, B. and Abdullah, M.A. (2014a). Heavy metal removal by oil palm empty fruit bunches (OPEFB) biosorbent. *Applied Mechanics and Materials*, **625**, 889 – 892.
- [19] Daneshfozoun, S., Nazir, M.S., Abdullah, B. and Abdullah, M.A. (2014b). Surface modification of celluloses extracted from oil palm empty fruit bunches for heavy metal sorption. *Chemical Engineering Transactions*, **37**, 679 – 684.
- [20] Danish, M., Ahmad, T., Hashim, R., Hafiz, M.R., Ghazali, A., Sulaiman, O., and Hiziroglu, S., (2015). Characterisation and adsorption kinetic study of surfactant treated oil palm (*Elaeis guineensis*) empty fruit bunches. *Desalination and Water Treatment*, **57**, 9474 – 9487.
- [21] Fatah, I.Y.A., Abdul Khalil, H.P.S., Hossain, M.S., Aziz, A.A., Davoudpour, Y., Dungani, R. and Bhat, A. (2014). Exploration of a chemo-mechanical technique for the isolation of nanofibrillated cellulosic fibre from oil palm empty fruit bunch as a reinforcing agent in composite materials. *Polymers*, **6**, 2611 – 2624.
- [22] Foo, K.Y. and Hameed, B.H. (2011). Preparation of oil palm (*Elaeis*) empty fruit bunch activated carbon by microwave-assisted KOH activation for the adsorption of methylene blue. *Desalination*, **275**, 302 – 305.
- [23] Food and Agriculture Organisation of the United Nations (FAO), (2018). Food and agriculture data. Retrieved on 4 December 2018. <http://www.fao.org/faostat/en/#home>
- [24] Funke, A. and Ziegler, F. (2010). Hydrothermal carbonisation of biomass: A summary and discussion of chemical mechanisms for process engineering. *Biofuels, Bioproducts and Biorefining*, **4**, 160 – 177.
- [25] Guo, J. and Lua, A.C. (2002). Textural and chemical characterisations of adsorbent prepared from palm shell by potassium hydroxide impregnation at different stages. *Journal of Colloid and Interface Science*, **254**, 227 – 233.
- [26] Hameed, B.H., Tan, I.A.W. and Ahmad, A.L. (2009). Preparation of oil palm empty fruit bunch-based activated carbon for removal of 2,4,6-trichlorophenol: Optimisation using response surface methodology. *Journal of Hazardous Materials*, **164**, 1316 – 1324.
- [27] Haron, M. J., Tiansih, M., Ibrahim, N. A., Kassim, A., & Wan Yunus, W. M. Z. (2009). Sorption of cu(II) by poly(hydroxamic acid) chelating exchanger prepared from poly(methyl acrylate) grafted oil palm empty fruit bunch (OPEFB). *BioResources*, **4**(4), 1305 – 1318.

- [28] Hashemipour, H., Baroutian, S., Jamshidi, E. and Abazari, A. (2009). Experimental study and artificial neural networks simulation of activated carbon synthesis in fluidized bed reactor. *International Journal of Chemical Reactor Engineering*, 7(1), 1 – 15.
- [29] Hidayu, A.R., Mohamad, N.F., Matali, S. and Shaifah, A.S.A.K. (2013). Characterisation of activated carbon prepared from oil palm empty fruit bunch using BET and FT-IR techniques. *Procedia Engineering*, 68, 379 – 384.
- [30] Ibrahim, M. N. M., Ngah, W. S. W., Norliyana, M. S., and Daud, W. R. W. (2009). Copper(II) biosorption on soda lignin from oil palm empty fruit bunches (EFB). *Clean - Soil, Air, Water*, 37(1), 80 – 85.
- [31] Idris, S.S., Rahman, N.A., Ismail, K., Alias, A.B., Rashid, Z.A. and Aris, M.J. (2010). Investigation on thermochemical behaviour of low rank Malaysian coal, oil palm biomass and their blends during pyrolysis via thermogravimetric analysis (TGA). *Bioresource Technology*, 101, 4584 – 4592.
- [32] Ibrahim, M. N. M., Ngah, W. S. W., Norliyana, M. S., Daud, W. R. W., Rafatullah, M., Sulaiman, O., and Hashim, R. (2010). A novel agricultural waste adsorbent for the removal of lead (II) ions from aqueous solutions. *Journal of Hazardous Materials*, 182(1-3), 377–385.
- [33] Jamari, S.S. and Howse, J.R. (2012). The effect of the hydrothermal carbonisation process on palm oil empty fruit bunch. *Biomass and Bioenergy*, 47, 82 – 90.
- [34] Johari, I. S., Yusof, N. A., Haron, M. J., and Mohd Nor, S. M. (2013). Preparation and characterization of poly(ethyl hydrazide) grafted oil palm empty fruit bunch for removal of Ni(II) ion in aqueous environment. *Polymers*, 5(3), 1056 – 1067.
- [35] Joseph, C.G., Quek, K.S., Daud, W.M.A.W. and Moh, P.Y. (2016). Physical activation of oil palm empty fruit bunch via CO₂ activation gas for CO₂ activation. *29th Symposium of Malaysia Chemical Engineers (SOMChE): Materials Science and Engineering*, 206, 012003.
- [36] Joseph, C.G., Wan Mohd, A.W.D., Quek, K.S. and Sanmygam, K. (2015). Parametric and adsorption kinetic studies of Reactive Black 5 removal from textile simulated wastewater using oil palm (*Elais guineensis*) empty fruit bunch. *Journal of Applied Sciences*, 15(8), 1103 – 1111.
- [37] Kadir, S.A.S.A., Matali, S., Mohamad, N.F. and Hidayu, A.R. (2014). Preparation of activated carbon from oil palm empty fruit bunch (EFB) by steam activation using response surface methodology. *International Journal of Materials Science and Applications*, 3(5), 159 – 163.
- [38] Kamaludin, N.H., Ghazali, A. and Daud, W.W. (2012). Potential of fines as reinforcing fibers in alkaline peroxide pulp of oil palm empty fruit bunch. *Bioresources*, 7(3), 3425 – 3438.
- [39] Kavitha, B., Jothimani, P. and Rajannan, G. (2013). Empty fruit bunch – A potential organic manure for agriculture. *International Journal of Science, Environment*, 2(5), 930 – 937.
- [40] Khosravihaftkhany, S., Morad, N., Teng, T. T., Abdullah, A. Z., & Norli, I. (2013). Biosorption of Pb(II) and Fe(III) from aqueous solutions using oil palm biomasses as adsorbents. *Water, Air, and Soil Pollution*, 224(3), 1455 – 1468.
- [41] Lee. T., Zubir, A.A., Jamil, F.M., Matsumoto, A. and Yeoh, F.Y. (2014). Combustion and pyrolysis of activated carbon fibre from oil palm empty fruit bunch fibre assisted through chemical activation with acid treatment. *Journal of Analytical and Applied Pyrolysis*, 110, 408 – 418.
- [42] Lenihan, P., Orozco, A., O'Neill, E., Ahmad, M.N.M., Rooney, D.W. and Walker, G.M. (2010). Dilute acid hydrolysis of lignocellulosic biomass. *Chemical Engineering Journal*, 156, 395 – 403.
- [43] Lillo-Rodenas, M.A., Cazorla-Amoros, D. and Linares-Solano, A. (2003). Understanding chemical reactions between carbons and NaOH and KOH: An insight into the chemical activation mechanism. *Carbon*, 41, 267 – 275.
- [44] Ma'an, F.A., Muyibi, S.A., & Amode, J.O. (2011). Optimization of activated carbon production from empty fruit bunch fibres in one-step steam pyrolysis for cadmium removal from aqueous solution. *Environmentalist*, 31, 349 – 357.

- [45] Mahjoub, R., Yatim, J.B.M. and Sam, A.R.M. (2013). A review of structural performance of oil palm empty fruit bunch fibre in polymer composites. *Advances in Materials Science and Engineering*, **2013**, 1 – 9.
- [46] Mohammed, M.A.A., Salmiaton, A., Wan Azlina, W.A.K.G. and Mohamad Amran, M.S. (2012). Gasification of oil palm empty fruit bunches: A characterisation and kinetic study. *Bioresource Technology*, **110**, 628 – 636.
- [47] Mubarak, N.M., Kundu, A., Sahu, J.N., Abdullah, E.C. and Jayakumar, N.S. (2014). Synthesis of palm oil empty fruit bunch magnetic pyrolytic char impregnation with FeCl₃ by microwave heating technique. *Biomass and Bioenergy*, **61**, 265 – 275.
- [48] Nassar, M. M., Ewida, K. T., Ebrahiem, E. E., Magdy, Y. H., & Mheaedi, M. H. (2004). Adsorption of iron and manganese using low cost materials as adsorbents. *Journal of Environmental Science and Health. Part A, Toxic/hazardous Substances & Environmental Engineering*, **39**(1), 421 – 434.
- [49] Nassar, M.M. and Magdy, Y.H. (1997). Removal of different basic dyes from aqueous solutions by adsorption on palm-fruit bunch particles. *Chemical Engineering Journal*, **66**, 223 – 226.
- [50] Kabbashi, N.A., Elwathig, M. and Jamil, I.N. (2011). Application of activated carbon from empty fruit bunch (EFB) for mercury [Hg(II)] removal from aqueous solution. *African Journal of Biotechnology*, **10**(81), 18768 – 18774.
- [51] Nwabanne, J. T., & Igbokwe, P. K. (2012). Mechanism of Copper (II) Removal from Aqueous Solution Using Activated Carbon Prepared from Different Agricultural Materials. *International Journal of Multidisciplinary Sciences and Engineering*, **3**(7), 46 – 52.
- [52] Omar, R., Idris, A., Yunus, R., Khalid, K. and Aida Isma, M.I. (2011). Characterisation of empty fruit bunch for microwave-assisted pyrolysis. *Fuel*, **90**, 1536 – 1544.
- [53] Ooi, C.H., Cheah, W.K., Sim, Y.L., Pung, S.Y. and Yeoh, F.Y. (2017). Conversion and characterisation of activated carbon derived from palm empty fruit bunch waste and its kinetic study on urea adsorption. *Journal of Environmental Management*, **197**, 199 – 205.
- [54] Parshetti, G.K., Hoekman, S.K. and Balasubramanian, R. (2013). Chemical, structural and combustion characteristics of carbonaceous products obtained by hydrothermal carbonisation of palm empty fruit bunches. *Bioresource Technology*, **135**, 683 – 689.
- [55] Puligundha, P., Oh, S. and Mok, C. (2016). Microwave-assisted pretreatment technologies for the conversion of lignocellulosic biomass to sugars and ethanol: a review. *Carbon Letters*, **17**(1), 1 – 10.
- [56] Rattanawong, O., Kaewsichan, L., Grisdanurak, N. and Yuasa, A. (2007). Sorption of oil emulsified in water on oil palm fibres. *Korean Journal of Chemical Engineering*, **24**(1), 67 – 71.
- [57] Rashidi, N.A. and Yusup, S., (2017). A review on recent technological advancement in the activated carbon production form oil palm waste. *Chemical Engineering Journal*, **314**, 277 – 290.
- [58] Rebitanim, N.Z., Ghani, W.A.W.A.K., Mahmoud, D.K., Rebitanim, N.A. and Salleh, A.A.A. (2012). Adsorption capacity of raw empty fruit bunch biomass onto methylene blue dye in aqueous solution. *Journal of Purity, Utility Reaction and Environment*, **1**, 45 – 60.
- [59] Sajab, M.S., Chia, C.H., Zakaria, S. and Khiew, P.S. (2013). Cationic and anionic modifications of oil palm empty fruit bunch fibres for the removal of dyes from aqueous solutions. *Bioresource Technology*, **128**, 571 – 577.
- [60] Sajab, M.S., Chia, C.H., Zakaria, S. and Sillanpaa, M. (2014). Removal of organic pollutants and decolourisation of bleaching effluents from pulp and paper mill by adsorption using chemically treated oil palm empty fruit bunch fibres. *Bioresources*, **9**(3), 4517 – 4527.
- [61] Sajab, M.S., Chia, C.H., Zakaria, S. and Sillanpaa, M. (2017). Adsorption of heavy metal ions on surface functionalised oil palm empty fruit bunch fibres: Single and Binary systems. *Sains Malaysiana*, **46**(1), 157 – 165.

- [62] Salamatinia, B., Kamaruddin, A. H., & Abdullah, A. Z. (2006). Removal of Zn and Cu from wastewater by sorption on Oil Palm tree-derived biomasses. *Journal of Applied Sciences*, 7(15), 2020 – 2027.
- [63] Samsuri, A.W., Sadegh-Zadeh, F. & She-Bardan, B.J. (2014). Characterisation of biochars produced from oil palm and rice husks and their adsorption capacities for heavy metal. *International Journal of Environmental Science and Technology*, 11, 967 – 976.
- [64] Samsuri, A.W., Sadegh-Zadeh, F. and She-Bardan, B.J. (2013). Adsorption of As(III) and As(V) by Fe coated biochars and biocars produced from empty fruit bunch and rice husk. *Journal of Environmental Chemical Engineering*, 1, 981 – 988.
- [65] Sari, N.A., Ishak, C.F. & Bakar, R.A. (2014). Characterisation of oil palm empty fruit bunch and rice husk biochars and their potential to adsorb arsenic and cadmium. *American Journal of Agricultural and Biological Sciences*, 9(3), 450 – 456.
- [66] Setiabudi, H.D., Jusoh, R., Suhaimi, S.F.R.M. and Masrur, S.F. (2016). Adsorption of methylene blue onto oil palm (*Elaeis guineensis*) leaves: Process optimisation, isotherm, kinetics and thermodynamic studies. *Journal of Taiwan Institute of Chemical Engineers*, 63, 363 – 370.
- [67] Shaarani, F.W. and Hameed, B.H. (2010). Batch adsorption of 2,4-dichlorophenol onto activated carbon agricultural waste. *Desalination*, 255, 159 – 164.
- [68] Shaarani, F.W. and Hameed, B.H. (2011). Ammonia-modified activated carbon for the adsorption of 2,4-dichlorophenol. *Chemical Engineering Journal*, 169, 180 – 185.
- [69] Shanmugarajah, B., Chew, I.M., Mubarak, N.M., Choong, T.S.Y., Yoo, C.K. and Tan, K.W. (2019). Valorisation of palm oil agro-waste into cellulose biosorbents for highly effective textile effluent remediation. *Journal of Cleaner Production*, 210, 697 – 709.
- [70] Shariff, A., Aziz, N.S.M. and Abdullah, N. (2014). Slow pyrolysis of oil palm empty fruit bunches for biochar production and characterisation. *Journal of Physical Science*, 25(2), 7 – 112.
- [71] Sukiran, M.A., Loh, S.K., Bakar, N.A. and Choo, Y.M. (2011). Production and characterisation of bi-char from the pyrolysis of empty fruit bunches. *American Journal of Applied Sciences*, 8(10), 984 – 988.
- [72] Tan, I.A.W. and Hameed, B.H. (2010). Adsorption isotherm, kinetics, thermodynamics and desorption studies of basic dye on activated carbon derived from oil palm empty fruit bunch. *Journal of Applied Science*, 10(21), 256 – 2571.
- [73] Tan, I.A.W., Ahmad, A.L. and Hameed, B.H. (2009). Adsorption isotherms, kinetics, thermodynamics and desorption studies of 2,4,6-trichlorophenol on oil palm empty fruit bunch-based activated carbon. *Journal of Hazardous Materials*, 164, 473 – 482.
- [74] Wahi, R., Ngaini, Z., & Jok, V. (2009). Removal of mercury, lead and copper from aqueous solution by activated carbon of palm oil empty fruit bunch. *World Applied Sciences Journal*, 5, 84 – 91.
- [75] Wirasnita, R., Hadibarata, T., Yusoff, A.R.M. & Lazim, Z.M. (2015). Preparation and characterisation of activated carbon form oil palm empty fruit bunch wastes using zinc chloride. *Jurnal Teknologi*, 74, 77 – 81.
- [76] Wirasnita, R., Hadibarate, T., Yusoff, A.R.M. and Yusop, Z. (2014). Removal of Bisphenol A from aqueous solution by activated carbon derived from oil palm empty fruit bunch. *Water, Air, and Soil Pollution*, 225, 2148 – 2159.
- [77] Yang, H., Yan, R., Chin, T., Liang, D.T., Chen, H. and Zheng, C. (2004). Thermogravimetric analysis–Fourier transform infrared analysis of palm oil waste pyrolysis. *Energy & Fuels*, 18, 1814 – 1821.
- [78] Yap, Y.L., Deraman, M., Jumali, M.H., Omar, R., Aziz, A.A., Abdelrahman, A.E., Tang, H.P., Tan, J.M., Muslimin, M. and Mahtar, M. (2005). Preparation and phenols adsorption property of porous carbon from oil palm empty fruit bunches. *Solid State Science and Technology*, 13(1-2), 170 – 178.

- [79] Yavari, S., Malakahmad, A., Sapari, N.B. and Yabari, S. (2017). Synthesis optimisation of oil palm empty fruit bunch and rice husks biochars for removal of imazapic and imazapyr herbicides. *Journal of Environmental Management*, **193**, 201 – 210.
- [80] Yusoff, S. (2006). Renewable energy from palm oil – innovation on effective utilization of waste. *Journal of Cleaner Production*, **14**, 87 – 93.
- [81] Zahrim, A.Y., Ricky, L.N.S., Shahril, Y., Rosalam, S., Nurmin, B., Harun, A.M. and Azreen, I. (2014). Partly decomposed empty fruit bunch fibre as potential adsorbent for ammonia-nitrogen from urban drainage water. *Proceedings of the International Civil and Infrastructure Engineering Conference (InCIEC)*. 28th September – 1st October, 2014. Kota Kinabalu, Sabah, Malaysia. pp 989 – 1001.
- [82] Zamani, S.A., Yunus, R., Samsuri, A.W., Mohd Salleh, M.A. and Asady, B. (2017). Removal of zinc from aqueous solution by optimized empty fruit bunches biochar as low cost adsorbent. *Bioorganic Chemistry and Application*, **7914714**, 1 – 9.